

Papéis Avulsos de Zoologia

Museu de Zoologia da Universidade de São Paulo

Volume 52(34):411-422, 2012

www.mz.usp.br/publicacoes
<http://portal.revistasusp.sibi.usp.br>
www.scielo.br/paz

ISSN impresso: 0031-1049

ISSN on-line: 1807-0205

CHANGE IN THE MEIOFAUNA COMMUNITY STRUCTURE OF SANDY BEACHES OF THE NUEVO GULF (CHUBUT, ARGENTINA)

CARLOS A. HARGUINTEGUY^{1,3}
M. NOELIA COFRE¹
CATALINA T. PASTOR DE WARD²

ABSTRACT

The composition and distribution of the benthic meiofauna assemblages of the Nuevo Gulf (Chubut, Argentina) are described in relation to abiotic variables. The meiofauna and sediment samples were collected in the intertidal zone of four sandy beaches with different anthropic disturbances in June 2005. The samples were obtained at 20 sampling sites using a 2.5 cm diameter core tube at a depth of 10 cm. A total of 13 meiofauna taxa were identified, with the meiofauna being primarily represented by nematodes, gastrotrichs, ciliates and polychaetes and the meiofauna abundances ranging from 1.5×10^3 to 6.5×10^3 ind. 10 cm^{-2} . Univariate (one-way ANOVA test) and multivariate (ANOSIM/MDS test) analyses showed clear dissimilarities in community structures between sites with anthropic effects and those in pristine condition, revealed by the significant differences were found between beaches near to and far way from a city with port activity. The meiofaunal assemblage varied in abundance and diversity, and these changes in the community structure may have been related to environmental gradients on the shore. The BIO-ENV analysis showed that the redox potential discontinuity depth might be the main factor in the spatial distribution of organisms.

KEY-WORDS: Meiofauna; Intertidal; Sandy sediment; Argentina.

INTRODUCTION

The interest of the benthic biologist in meiofauna increases started in the early 1980s. Probably the main obstacles of benthic meiofauna are its small size, together with difficulties in isolating the meiofauna from the sediments and the identification of species belonging to different taxa (Austen *et al.*, 1994).

Meiofaunal organisms play an important ecological role in the aquatic ecosystem and are well suited for environmental impact assessment studies. They have short generation times, continuous reproduction, and in situ direct development and, therefore, their potential for rapid response to environmental change is high (Giere, 1993; Fraschetti *et al.*, 2006; Gyedu-Ababio & Baird, 2006). The marine meiofauna is

1. Instituto Multidisciplinario de Biología Vegetal (IMBIV) – Facultad de Ciencias Exactas, Físicas y Naturales (FCEyN) – Universidad Nacional de Córdoba (UNC), Avda. Vélez Sarsfield 1611, Córdoba, Argentina.

2. Centro Nacional Patagónico (CENPAT). Bv. Brown (9120), Pto. Madryn, Chubut, Argentina.

3. E-mail corresponding author: c_harguinteguy@com.uncor.edu

often a very useful tool for biological monitoring since the community structure may be sensitive to both natural and anthropogenic environmental disturbance (Mirto & Danovaro, 2004; Gyedu-Ababio & Baird, 2006; Moreno *et al.*, 2008). In addition, the beaches may function as natural filters responsible for the remineralization of substances, which then return to the sea as nutrients (Coull & Chandler, 2001). The interstitial system of the beaches and marshes, in particular the system protected by muddy sediments, is formed by long and intricate food chains of bacteria, unicellular algae and meiofauna at the first levels. Therefore, biological systems are dependent on the productivity of coastal areas (Higgins & Thiel, 1988; Leguerrier *et al.*, 2003).

The growth and diversity of the meiofauna may be stimulated by feeding on bacteria, which could increase the recycling of nutrients into the ecosystem and thereby be expected to have a greater productivity (de Wit *et al.*, 2001; De Troch *et al.*, 2006). Moreover, the meiofauna can provide food for higher trophic levels, such as fish and marine invertebrates (Leduc & Probert, 2009). The spatial patterns of the structure of the meiofaunal community in sandy beaches of marine ecosystems may be associated with different environmental variables. Related to this, the sediment granulometry (Gómez Noguera & Hendrickx, 1997; Barnes *et al.*, 2008), the organic matter source in coastal sediments (Danovaro *et al.*, 2002; Flach *et al.*, 2002; Moreno *et al.*, 2008; Ingels *et al.*, 2009; Puscedu *et al.*, 2009), and oxic and anoxic conditions in the interstitial pore space (Mirto *et al.*, 2000; Sutherland *et al.*, 2007) have a fundamental role in the richness and abundance of the benthic meiofauna.

The criteria in the study of benthic meiofauna were established by Giere (1993) and these concepts have been recently applied in Latin America. Various studies were carried out on the South American coasts, in the Magallanes' strait (Chen *et al.*, 1999), on the Brazilian coast (Netto *et al.*, 2005; Netto *et al.*, 2009) and on the Chilean coast (Neira *et al.*, 2001; Lee *et al.*, 2006; Veit-Köhler *et al.*, 2009). However, none of the studies took place on the Argentinean coasts.

In this paper, we analyze the structure of the meiofauna assemblages in the intertidal sandy sediment of the Nuevo Gulf (Chubut, Argentina), over increasing distances from the beach of Puerto Madryn to the mouth of the gulf. The objective of this work was to evaluate the meiofaunal community structure and their relation with measuring the abiotic variables, in order to determine if the number of taxa and abundances of meiofauna change along of different beaches with anthropic effects.

MATERIAL AND METHODS

Study area

The study area is located on the southern coasts of Nuevo Gulf (42°43'S, 65°02'W to 42°50'S, 64°52'W), and in the southern part of Península Valdés on the Atlantic coast of South America (Chubut province, Argentina) (Fig. 1). The physical characteristics are similar to dissipative beaches with fine sediment. The coastal area is characterized by rock outcrops of easily eroded materials, such as marine sediments, sandstones, tuffs and silt-stones (Haller, 1981).

The tidal current is the most important water movement and the regime is semi-diurnal, ranging in altitude from 4 to 7 m. The Nuevo Gulf area is situated in the arid Patagonian region, characterized by low and irregular rainfall regimes (173 mm per year) and strong, frequent winds. The average surface water temperature is 13.5°C, ranging from 9.8°C (August-September) to 16.5°C (February). The spring phytoplankton bloom starts in early October with a second smaller bloom occurring in late February. The mean chlorophyll *a* concentration is 0.5 mg m⁻³, the primary production being limited by nitrate availability (Charpy & Charpy, 1977; Charpy-Roubaud *et al.*, 1982).

The city of Puerto Madryn, on the west coast of the Nuevo gulf, experiences notable changes in the characteristics of its surrounding waters due to sewage, storm-water and industrial discharges on its beaches. However, according to Esteves & De Vido de Mattio (1980), the variation in salinity in Nueva Gulf may be slower when compared to surrounding areas (the South). Therefore, in the vicinity of this bay, the movement of the water body may not participate in the general movement of the gulf, which depends mainly on the instantaneous direction and intensity of winds and tides (Krepper & Rivas, 1979).

Sampling and sample processing

During June 2005, a systematic sampling was conducted on mesolitoral, and 20 samples were taken in areas with the presence of tubeworm polychaeta communities (Maldanidae family). The sampling sites were located at four sandy beaches separated by a rocky area called "restinga": At Nueva beach 11 sites were selected, and at Kaiser beach, Paraná beach and Cerro Avanzado beach three sites were selected for each beach (Fig. 1). Three replicates at each site were taken manually using a 2.5 cm diameter core tube at a depth of 10 cm to analyze the meiofauna.

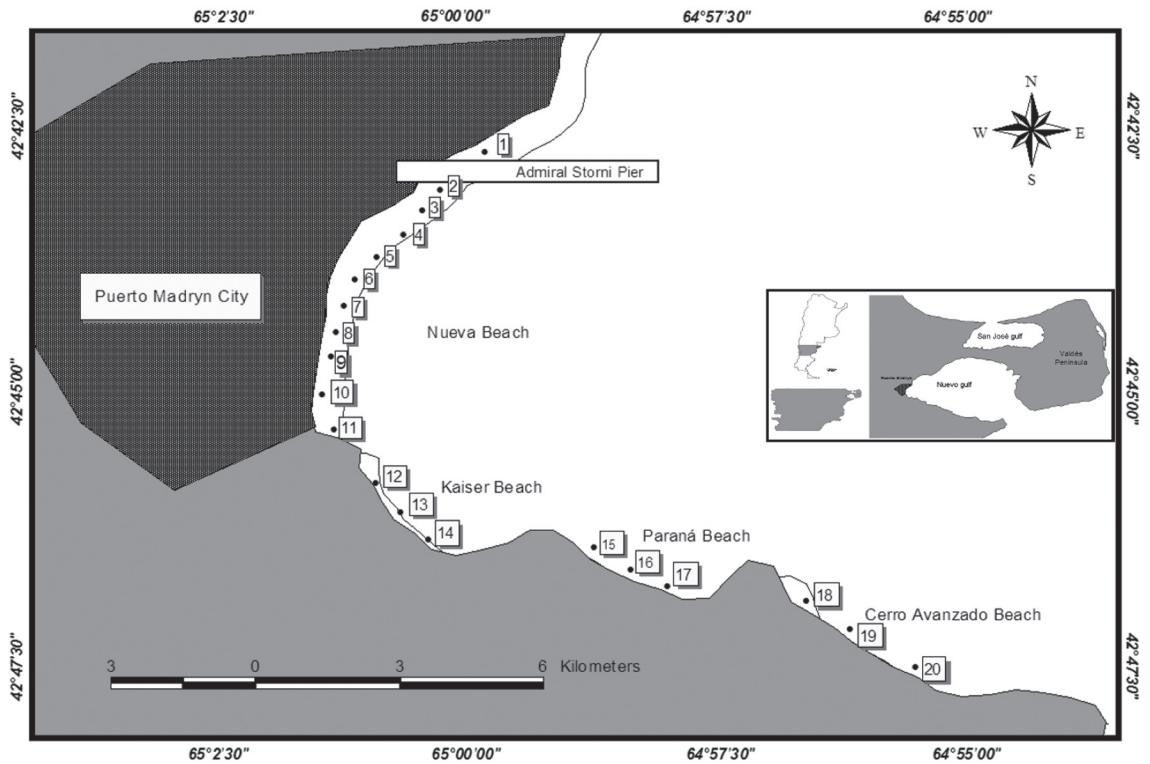


FIGURE 1: Location of study area in sandy beaches of Nuevo Gulf, showing the sampling sites (●).

All sampling units were preserved in 4% buffered formalin and stained with Rose Bengal. Later, these sampling units were elutriated of larger sand particles using a shake and decant procedure (Cross & Curran, 2000) and sieved through 0.50 and 0.05 mm mesh sizes. The content of the 0.05 mm sieve was recovered and preserved in the fixative Ditlevsen (1911). Then, the fauna were identified to higher taxa and counted under a stereomicroscope (Higgins & Thiel, 1988). The meiofaunal density was standardized to individuals per 10 cm².

Three replicates at each site were taken manually to analyze the sediment grain size, using a 2.5 cm diameter core tube at a depth of 5 cm. The mean grain size (MGS) was calculated, by sieving dried samples according to the International Test Sieve Standard R565, and the asymmetry and sorting coefficient were determined (Giere *et al.*, 1988). The percentage of silt/clay in the sediment was obtained by wet sieving using a 63 µm sieve to separate the fine and sand fractions, which were then dried at 80°C and weighed. The redox potential discontinuity depth (RPD in cm) was estimated visually as the depth at which the sediment color turned from brown to black. Sediment permeability was determined according to Jaramillo *et al.* (1993) and sediment temperatures (using a hand-held mercury thermometer) were also recorded at each site.

Data analyses

Univariate measures of the meiofauna taxa for each site were calculated, including number of taxa (S), total abundance (N), the Shannon-Wiener diversity indices (H') (calculated using Loge), Margalef's species richness (d) and evenness (Pielou's J) (Gobin & Warwick, 2006). The significance of differences in univariate measures between sites was tested using a one-way ANOVA. When necessary, data were fourth-root transformed to approximate normality. Tukey's HSD multiple comparison tests were used when significant differences were detected ($p < 0.05$). The similarity of meiofauna among sites was determined by cluster and non-metric multidimensional scaling ordinations (MDS) on fourth-root transformed data, using the Bray-Curtis similarity index. Formal significance tests for differences in the meiofauna community structure between sites were performed using the one-way ANOSIM test. Species contributing to dissimilarities between habitats were investigated using the similarities percentages procedure (SIMPER). For further details of the methods used, see Warwick & Clarke (1993) and Netto *et al.* (1999a).

Abiotic data were converted to approximate normality using a fourth-root transformation and ordinated using a correlation-based principal component

analysis (PCA). Differences in abiotic variables between sites were then tested using a one-way ANOVA on fourth-root transformed data. In order to examine the relationships between benthic community patterns and the environmental structure of the sites, a Spearman rank correlation (*p*) was computed between the Bray-Curtis similarity faunal matrices and the Euclidean distance matrix derived from abiotic data in order to analyze the relationships between abiotic

data and the benthic community structure. The relationships between the multivariate community structure and combinations of abiotic variables were then analysed using the BIO-ENV procedure (Warwick & Clarke, 1993; Netto *et al.*, 1999a) to define suites of variables which best explained the faunistic structure. All analyses were carried out with PRIMER v6 (Plymouth Routine In Multivariate Ecological Research) (Clarke *et al.*, 2005).

TABLE 1: Mean abundance (ind. 10 cm⁻² ± SD) and total percentage of meiofauna taxa in sampling sites of Nuevo Gulf.

Taxa	1	2	3	4	5	6	7
Nematoda	5937.33 ± 2521.65	6101.67 ± 304.51	6147.33 ± 1476.59	6027.33 ± 2681.67	4553.33 ± 1669.70	5360.67 ± 869.23	2808.67 ± 303.01
Gastrotricha	330.67 ± 174.91	90.67 ± 8.08	27.33 ± 24.17	45.33 ± 8.62	26.33 ± 24.50	56.67 ± 19.86	137.33 ± 95.26
Ciliophora	105.67 ± 94.88	24.67 ± 8.02	10.33 ± 9.24	20.67 ± 22.85	24.00 ± 20.42	18.67 ± 13.58	115.33 ± 36.30
Polychaeta	14.33 ± 11.37	1.00 ± 1.73	0.67 ± 1.15	0.67 ± 1.15	4.33 ± 2.08	0.67 ± 1.15	5.33 ± 2.52
Turbellaria	62.67 ± 41.19	25.33 ± 8.39	28.33 ± 24.21	105.67 ± 26.08	12.33 ± 2.89	14.67 ± 20.21	40.00 ± 17.06
Copepoda	5.33 ± 2.52	6.33 ± 2.89	5.33 ± 4.93	1.67 ± 2.89	1.00 ± 1.73	2.67 ± 2.52	1.00 ± 1.73
Nemertina	34.67 ± 12.34	—	—	—	—	—	—
Cumacea	0.67 ± 1.15	—	—	—	1.00 ± 1.73	—	0.67 ± 1.15
Anhipoda	—	—	1.00 ± 1.73	0.67 ± 1.15	—	—	—
Foraminifera	—	—	—	—	—	—	—
Halacaroida	—	—	1.00 ± 1.73	—	—	—	—
Ostracoda	—	0.67 ± 1.15	—	—	—	—	—
Oligochaeta	0.67 ± 1.15	—	—	0.67 ± 1.15	—	—	—
Taxa	8	9	10	11	12	13	14
Nematoda	2264.00 ± 201.48	1719.67 ± 411.62	2914.67 ± 642.32	3676.33 ± 897.14	3136.67 ± 457.73	2814.00 ± 1343.25	2945.00 ± 1544.21
Gastrotricha	628.33 ± 522.99	246.67 ± 49.86	104.00 ± 77.49	92.00 ± 81.50	23.33 ± 10.41	144.33 ± 199.51	236.00 ± 322.39
Ciliophora	229.33 ± 52.27	611.33 ± 353.63	57.67 ± 78.36	20.00 ± 15.1	4.33 ± 3.21	4.00 ± 3.46	4.00 ± 1.73
Polychaeta	196 ± 112.01	5.00 ± 3.46	1.00 ± 1.73	9.67 ± 11.93	0.67 ± 1.15	575.33 ± 642.09	—
Turbellaria	8.67 ± 6.51	11.33 ± 12.74	45.33 ± 12.5	1.67 ± 1.53	1.67 ± 1.53	2.67 ± 4.62	9.67 ± 7.02
Copepoda	5.00 ± 6.24	3.67 ± 4.04	2.00 ± 0.00	—	1.00 ± 1.73	0.67 ± 1.15	2.67 ± 4.62
Nemertina	—	—	—	—	—	—	0.67 ± 1.15
Cumacea	—	0.67 ± 1.15	1.33 ± 1.15	—	2.33 ± 0.58	1.33 ± 1.15	0.67 ± 1.15
Anhipoda	—	—	—	0.67 ± 1.15	1.67 ± 1.53	—	—
Foraminifera	—	—	—	—	—	—	0.67 ± 1.15
Halacaroida	—	—	—	—	—	0.67 ± 1.15	—
Ostracoda	—	0.67 ± 1.15	—	—	—	0.67 ± 1.15	—
Oligochaeta	—	—	—	—	—	—	—
Taxa	15	16	17	18	19	20	Total %
Nematoda	1674.33 ± 709.89	3535.33 ± 1551.74	2874.67 ± 1704.39	3388.67 ± 2228.08	1558.67 ± 471.93	2345.00 ± 883.73	93.19
Gastrotricha	47.67 ± 45.63	209.00 ± 323.99	66.33 ± 53.29	5.33 ± 2.52	9.33 ± 2.89	49.33 ± 25.66	3.34
Ciliophora	16.33 ± 19.09	8.00 ± 10.39	21.00 ± 14.80	4.33 ± 1.15	2.67 ± 0.58	10.00 ± 10.54	1.70
Polychaeta	0.67 ± 1.15	0.67 ± 1.15	0.67 ± 1.15	1.00 ± 1.73	12.67 ± 21.94	2.00 ± 3.46	1.08
Turbellaria	3.67 ± 4.04	2.67 ± 0.58	13.67 ± 16.5	—	—	11.33 ± 12.74	0.52
Copepoda	5.00 ± 3.46	3.00 ± 1.73	9.33 ± 5.13	3.33 ± 1.53	2.33 ± 0.58	2.33 ± 2.52	0.08
Nemertina	—	0.67 ± 1.15	—	0.67 ± 1.15	—	—	0.05
Cumacea	—	0.67 ± 1.15	1.33 ± 1.15	—	1.67 ± 1.53	1.67 ± 2.89	0.02
Anhipoda	0.67 ± 1.15	—	—	—	—	—	0.01
Foraminifera	—	—	—	—	—	—	< 0.01
Halacaroida	—	—	—	—	—	—	< 0.01
Ostracoda	—	—	—	—	—	—	< 0.01
Oligochaeta	—	—	1.67 ± 2.89	—	—	—	< 0.01

RESULTS

The meiofauna

The number of taxa, abundances and the total percentage of meiofauna in sampling sites of Nuevo Gulf are shown in Table 1. The number of taxa ranged from 6 to 9 per site, with abundance per site ranging from 1.5×10^3 to 6.5×10^3 ind. 10 cm^{-2} , and the mean abundance being 3.8×10^3 ind. 10 cm^{-2} . Nematodes were numerically dominant at all sites, accounting for more than 90% of the total meiofauna. Gastrotrichs (3%) were also present, but at lower abundances. Interstitial ciliates, polychaetes, turbellarians, nemertines, copepods, cumaceans, amphipods, halacarids, ostracods and oligochaetes were also poorly represented. At sites 8 and 9, located at the central region of Nueva beach, the nematodes did not exceed 68% of all analyzed taxa, with the gastrotrichs (19%), polychaetes (7%) and ciliates (23%) gaining in importance.

The number of taxa and abundance of meiofauna in beaches near to and far away from a city of Nuevo Gulf are shown in Table 2. Foraminifera was absent in beaches near the city. The abundance of nematodes, gastrotrichs, ciliates, turbellarians, nemertines and ostracods were higher in beaches near the city, except for polychaetes, copepods, cumaceans, amphipods and oligochaetes, which were higher in beaches far away.

The univariate measures derived from the meiofauna data in sampling sites are shown in Fig. 2. Significant differences were found using Tukey's HSD multiple comparison tests ($p < 0.05$). The total abundance (N) differed significantly ($p = 0.0004$) between the sites, with the lower values occurring at the Kaiser, Paraná, and Cerro Avanzado beaches. The diversity (H') differed significantly ($p < 0.0001$) between the sites, and the highest values were found at sites 8 and 9 of the Nueva beach, but neither the number of taxa (S) nor the evenness (J) differed significantly between the sites ($p = 0.4971$ and $p = 0.5834$, respectively).

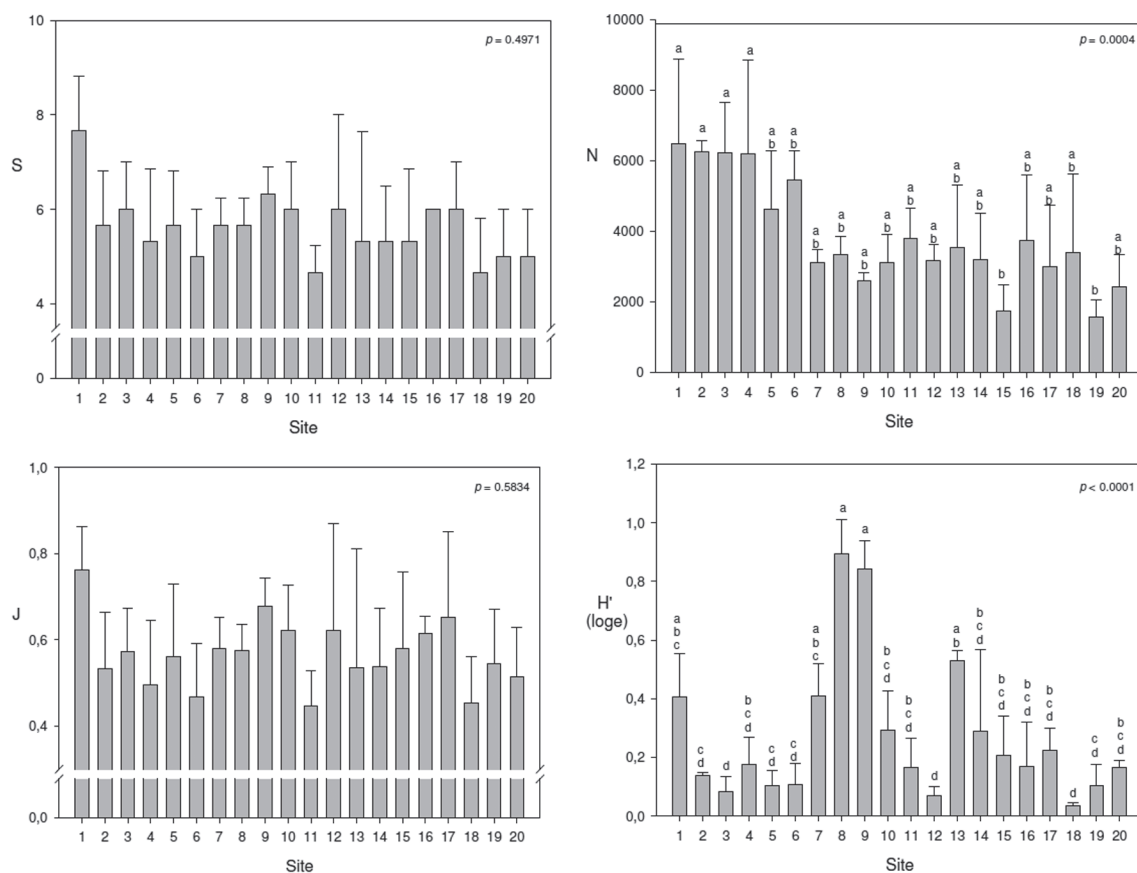


FIGURE 2: Means and standard deviation for univariate measures of the meiofauna community structure at different sites of Nuevo Gulf for number of taxa (S), abundance (N), Shannon-Wiener diversity (H') and evenness (Pielou's J). Analysis of variance (ANOVA). Note: Bars with the same letters are not significantly different (Tukey test, $p < 0.05$).

TABLE 2: Mean abundance (ind. 10 cm⁻² ± SD) of meiofauna taxa in beaches near to and far away from a city of Nuevo gulf.

Taxa	Near to	Far away
Nematoda	4319.18 ± 2002.37	2696.93 ± 1303.94
Gastrottricha	162.30 ± 226.73	87.85 ± 162.95
Ciliophora	112.52 ± 197.13	8.30 ± 10.09
Polychaeta	21.70 ± 62.88	65.96 ± 255.83
Turbellaria	32.36 ± 33.51	5.04 ± 8.04
Copepoda	3.09 ± 3.39	3.30 ± 3.47
Nemertina	3.15 ± 10.58	0.22 ± 0.64
Cumacea	0.39 ± 0.86	1.07 ± 1.36
Anphipoda	0.21 ± 0.70	0.26 ± 0.76
Foraminifera	—	0.07 ± 0.38
Halacaroida	0.09 ± 0.52	0.07 ± 0.38
Ostracoda	0.12 ± 0.48	0.07 ± 0.38
Oligochaeta	0.12 ± 0.48	0.19 ± 0.96

The simulated distribution of the test statistic R in the ANOSIM analysis of the meiofauna community structure between beaches near to and far away from a city of Nuevo Gulf was realized *a priori*. Results of the global ANOSIM tests ($R = 0.412$; $p < 0.1$) confirmed that the structure of the meiofauna community was different between beaches near to and far away from the city

The MDS ordination of the meiofauna data from Nuevo Gulf sites (Fig. 3) indicated that the beaches near to the city of Nuevo Gulf showed differences from those far away from the Kaiser, Paraná and Cerro Avanzado beaches. The low stress factor of 0.15 indicates a good MDS ordination with no real prospect of a misleading interpretation. The analyses revealed clear differences in the community structures between the sites of the gulf, particularly between the Nueva beach and the other beaches.

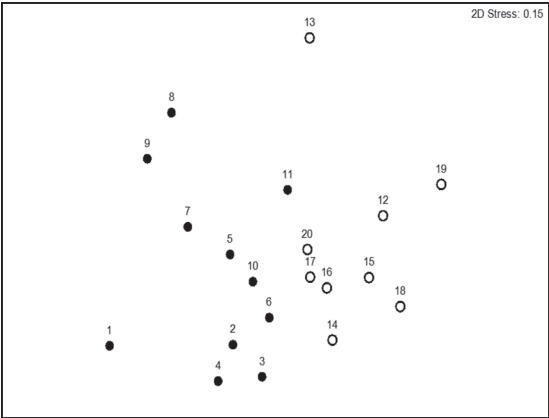


FIGURE 3: MDS ordinations from fourth root transformed abundances of meiofauna data in the beaches near to (●) and far away from (○) a city of Nuevo Gulf.

TABLE 3: The mean univariate measures of abundance of the meiofauna data collected in the beaches near to and far away from a city of Nuevo Gulf. Values are the means ± SD. Analysis of variance (ANOVA) and Tukey test ($p < 0.05$).

Univariate measures	S	N	J	H' (loge)
Near to-Far away				
Near to	5.8 ± 1.1	4654.7 ± 1845.6	0.57 ± 0.13	0.33 ± 0.29
Far away	5.4 ± 1.2	2868.2 ± 1397.9	0.56 ± 0.16	0.20 ± 0.17
F-value	ns	19,4	ns	4,04
p	0.2170	< 0.0001	0.7116	0.0491

Note: ns = not-significant; $p < 0.05$ = is significant; S = number of taxa; N = total abundance; H' = Shannon-Wiener diversity indices calculated using Loge; J = Pielou's evenness.

The SIMPER results demonstrated average similarities of 71.9% between beaches near to and far away from the city of Nuevo gulf. The taxa contributing most to the similarity term were Turbellaria (19.7%), Ciliophora (19.0%), Nematoda (17.5%) and Polychaeta (15.7%).

Abiotic variables

All abiotic variables, except for asymmetry and sorting coefficients, varied significantly between the sites (Table 4i) and showed a gradual change from port to pristine areas. Sediments ranged from fine to silt-clay sand and from moderately well to very well sorted. The values of permeability in the sediment, around 3.0 ± 0.0 cm, were relatively low near to the Admiral Storni pier and showed a significant increase in the Kaiser and Paraná beaches (sites 11, 15, 16, 17), although a decrease was observed in the sediment permeability in the Cerro Avanzado beach. The temperature, around $9.0 \pm 1.0^\circ\text{C}$ (site 8), were significantly lower in the central area of Nuevo Gulf, and the redox potential discontinuity depth was significantly higher in both the Paraná and Cerro Avanzado beaches (site 15: 12.00 ± 0.00 cm and site 19: 12.33 ± 1.53 cm, respectively). In fact, the redox potential discontinuity showed an increasing trend along the entire sampled intertidal area, where the thinnest layer was observed next to the Admiral Storni pier and the central area of Nueva beach (site 4). The oxygenated layer values revealed a significant increase from the Kaiser beach to the Cerro Avanzado beach, and the mean grain size was significantly higher near to Admiral Storni pier (site 1: $191.00 \pm 19.28 \mu\text{m}$) and in the central area of the Nueva beach next to a rocky area (site 9: $180.93 \pm 0.85 \mu\text{m}$, site 10: $181.73 \pm 1.50 \mu\text{m}$ and site 11: $189.60 \pm 9.53 \mu\text{m}$) as well as at Cerro

TABLE 4: Mean abiotic variables in sediment collected at the different sampling sites (i) and in the beaches near to and far away from a city of Nuevo gulf (ii). Values are the means \pm SD. Analysis of variance (ANOVA) and Tukey test ($\alpha = 0.05$) ($n = 3$ samples per site).

i) Abiotic variables	Sk	So	Permeability (cm)	Temperature (°C)	RPD (cm)	MGS (μ m)	Fine fract. %
Site							
1	0.77 \pm 0.35	1.51 \pm 0.49	1.7 \pm 0.6 a	9.3 \pm 1.2 ab	5.33 \pm 1.04 ab	191.00 \pm 19.28 b	2.50 \pm 0.07 bc
2	1.00 \pm 0.02	1.15 \pm 0.02	2.0 \pm 0.5 ab	11.0 \pm 1.0 c	8.17 \pm 1.04 bcde	166.23 \pm 5.02 ab	3.02 \pm 0.50 c
3	0.99 \pm 0.01	1.14 \pm 0.01	1.8 \pm 0.3 ab	11.0 \pm 1.0 c	7.33 \pm 1.04 abcd	171.67 \pm 7.01 ab	2.18 \pm 0.44 abc
4	1.01 \pm 0.01	1.17 \pm 0.02	2.3 \pm 0.3 ab	11.0 \pm 1.0 c	4.07 \pm 1.4 a	154.17 \pm 14.43 ab	2.57 \pm 0.52 bc
5	1.00 \pm 0.03	1.18 \pm 0.01	2.2 \pm 0.3 ab	11.0 \pm 1.0 c	6.00 \pm 1.00 abc	143.77 \pm 18.75 ab	2.13 \pm 0.24 abc
6	1.05 \pm 0.06	1.19 \pm 0.04	2.3 \pm 0.3 ab	11.0 \pm 1.0 c	7.67 \pm 1.26 bcde	158.33 \pm 7.22 ab	2.51 \pm 0.51 bc
7	1.06 \pm 0.10	1.19 \pm 0.06	2.3 \pm 0.3 ab	11.2 \pm 0.3 c	8.50 \pm 0.50 bcde	161.80 \pm 5.05 ab	1.99 \pm 0.27 abc
8	0.98 \pm 0.01	1.13 \pm 0.00	2.8 \pm 0.3 ab	9.0 \pm 1.0 a	8.17 \pm 1.44 bcde	175.73 \pm 1.59 ab	1.43 \pm 0.30 abc
9	0.99 \pm 0.00	1.12 \pm 0.01	2.2 \pm 0.3 ab	9.7 \pm 0.6 abc	6.17 \pm 0.76 abc	180.93 \pm 0.85 b	1.79 \pm 0.17 abc
10	0.99 \pm 0.00	1.12 \pm 0.00	2.3 \pm 0.6 ab	10.3 \pm 0.6 abc	9.67 \pm 2.25 cde	181.73 \pm 1.50 b	1.65 \pm 0.18 abc
11	0.98 \pm 0.00	1.14 \pm 0.01	3.0 \pm 0.0 b	9.7 \pm 0.6 abc	6.67 \pm 0.76 abcd	189.60 \pm 9.53 b	2.23 \pm 1.24 abc
12	1.00 \pm 0.01	1.18 \pm 0.03	2.0 \pm 0.0 ab	10.7 \pm 0.6 bc	10.83 \pm 2.57 de	162.60 \pm 6.50 ab	1.10 \pm 0.72 a
13	0.98 \pm 0.02	1.19 \pm 0.00	3.0 \pm 1.0 ab	11.0 \pm 1.0 c	7.67 \pm 0.58 bcde	164.87 \pm 57.53 ab	2.25 \pm 0.28 abc
14	1.02 \pm 0.01	1.19 \pm 0.01	2.3 \pm 0.6 ab	11.0 \pm 1.0 c	8.50 \pm 2.50 bcde	144.27 \pm 5.51 ab	2.04 \pm 0.27 abc
15	1.00 \pm 0.02	1.19 \pm 0.00	3.0 \pm 0.0 b	11.0 \pm 1.0 abc	12.00 \pm 0.00 e	128.00 \pm 2.60 a	1.58 \pm 0.06 abc
16	1.01 \pm 0.02	1.18 \pm 0.02	3.0 \pm 0.0 b	10.0 \pm 0.0 abc	10.83 \pm 0.29 de	149.13 \pm 12.40 ab	1.35 \pm 0.05 ab
17	0.99 \pm 0.03	1.19 \pm 0.01	3.0 \pm 1.0 ab	10.0 \pm 0.0 abc	9.00 \pm 0.00 bcde	143.60 \pm 31.50 ab	1.65 \pm 0.29 abc
18	0.98 \pm 0.00	1.15 \pm 0.02	1.8 \pm 0.3 ab	11.0 \pm 1.0 c	6.83 \pm 0.76 abcd	173.03 \pm 7.16 ab	2.11 \pm 0.99 abc
19	1.00 \pm 0.02	1.16 \pm 0.03	2.0 \pm 0.0 ab	11.0 \pm 1.0 c	12.33 \pm 1.53 e	161.47 \pm 9.05 ab	2.29 \pm 0.40 abc
20	0.93 \pm 0.11	1.21 \pm 0.10	1.8 \pm 0.3 ab	11.0 \pm 1.0 c	7.00 \pm 2.00 abcd	185.77 \pm 19.30 b	1.88 \pm 0.57 abc
F-value	ns	ns	3.19	5.58	7.6	2.99	2.7
p	0.2925	0.0844	0.001	< 0.0001	< 0.0001	0.0018	0.0041
ii) Abiotic variables	Sk	So	Permeability (cm)	Temperature (°C)	RPD (cm)	MGS (μ m)	Fine fract. %
Near to-Far away							
Near to	0.98 \pm 0.12	1.19 \pm 0.16	2.3 \pm 0.5	10.4 \pm 0.9	7.07 \pm 1.85	170.45 \pm 16.93	2.18 \pm 0.61
Far away	0.99 \pm 0.04	1.18 \pm 0.04	2.4 \pm 0.7	10.6 \pm 0.5	9.44 \pm 2.38	156.97 \pm 25.88	1.81 \pm 0.57
F-value	ns	ns	1.08	1.68	18.02	6.45	6.11
p	0.7581	0.9482	0.3024	0.2003	0.0001	0.0138	0.0164

Note: ns = not-significant; $p < 0.05$ = is significant; different letters show differences between means; Sk = asymmetry coefficients; So = sorting coefficients; RPD = redox potential discontinuity depth; MGS = mean grain size; Fine fract. % = silt/clay percentage.

Avanzado beach (site 20: 185.77 \pm 19.30 μ m). The fine fraction percentage was significantly higher near to Admiral Storni pier (site 2: 3.02 \pm 0.50%), with a marked decrease occurring towards the restinga area of Nueva beach. In addition, the fine fraction percentage showed an increase in both the Kaiser and Cerro Avanzado beaches. All the abiotic variables, except for the asymmetry and mixture coefficients, permeability and temperature, varied significantly between beaches near to and far away from the city of Nuevo Gulf (Table 4ii). The redox potential discontinuity depth was significantly higher in beaches far away from the city (9.44 \pm 2.38 cm) whereas mean grain size was significantly higher in beaches near to the city (170.45 \pm 16.93 μ m), with the fine fraction percentage also being significantly higher at this location (2.18 \pm 0.61%).

The ordination by PCA (Fig. 4) of abiotic data showed the clear site distinction of the gulf. The first two components explained 72.9% of the data variance (PC1 = 46.7%, PC2 = 26.2%). Sites 9, 10 and 11 in beaches near to the a city were associated with the mean grain size, and the sites 12, 15, 16 and 17 in beaches far away from the city were mainly associated to the redox potential discontinuity depth.

The results of Spearman correlations analyses showed that the meiofauna was significantly correlated with abiotic variables ($p < 0.001$). For a Pearson correlation analysis of all the measured abiotic variables, the highest correlation was 0.65. Thus, all variables, except for asymmetry and sorting coefficients, were used in the BIO-ENV analyses to define sets of abiotic variables, which showed the highest correlation values with the meiofauna data. The results for

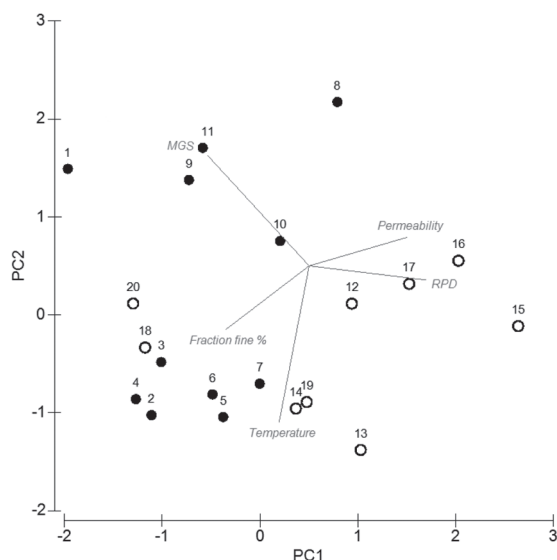


FIGURE 4: Principal component analysis ordinations from fourth root transformed data of abiotic variables in the beaches near to (●) and far away from (○) the city of Nuevo Gulf. RPD: redox potential discontinuity depth; MGS: mean grain size.

the meiofauna data revealed correlation values lower than 0.45, which therefore were not well explained by the measured abiotic variables. However, the results of MDS and PCA ordinations showed similar patterns, and the meiofauna may have presented a similar pattern mainly due to the redox potential discontinuity depth.

DISCUSSION

Meiofaunal characteristics

The meiofaunal abundance recorded in the beaches of the Nuevo Gulf (1.5×10^3 and 6.5×10^3 ind. 10 cm^{-2}) was found within the range elsewhere for template beaches with fine sediment. This abundance is approximately 10^3 individuals per 10 cm^2 , according to Gómez Noguera & Hendrickx (1997), Dittmann (2000) and Veit-Köhler *et al.* (2009). However, it is higher than for Mediterranean Sea beaches with fine sediment (Mirto *et al.*, 2000; Flach *et al.*, 2002; Moreno *et al.*, 2008). In this study area, the numbers of taxa of meiofauna (6 to 9 taxa) were lower than the values obtained by Rodríguez *et al.* (2003) of sandy beaches in northern Spain, *i.e.* between 8 and 14 taxa per beach. We observed that the most abundant taxa were nematodes, gastrotrichs and ciliates, in contrast with the findings of other studies on coastal environments, where Dittmann (2000), Rodríguez *et al.* (2003) and Sajan *et al.*

(2010) demonstrated that copepod taxa was second in the order of dominance. The number of taxa and abundance of meiofauna were higher in beaches near to than far away from a city of Nuevo Gulf (Tabla 3). On the other hand, the polychaete abundance was higher in beaches far away from than near to this city. These results were in agreement with Frascchetti *et al.* (2006) who demonstrated that the meiofauna community structure change by anthropic effects and in particular the polychaetes abundance is negatively affected in zone with sewage discharges.

Relationships between meiofauna and sediment characteristics

Sediments from the beaches of Nuevo Gulf were symmetric, homogeneously sorted and with fine sand, revealing a low energy environment. These sediments are characteristic of dissipative beaches, which have gentler slopes and often have fine sand (Yamanaka *et al.*, 2010). Both the abundance and diversity of the benthic meiofauna in this area reflected the degree of disturbance to which this community was subjected, with the meiofauna being characterized by its high density and relatively low diversity in the beaches of Nuevo Gulf. However, these values do not concur with those of Raffaelli & Hawkins (1996) or Yamanaka *et al.* (2010), who concluded that more dissipative beaches with gentler slopes and finer particle sizes often support a higher number of species and a greater abundance.

The meiofauna community structures clearly showed a pattern that was related to differences in the physical environments of the beaches of the sand flats of the Nuevo Gulf, with the highest abundance being observed near to the Admiral Storni pier and in the central areas of Nueva beach. This area was characterized by high percentages of fine fraction, low permeability of the sediment and a thin oxygenated layer. The Admiral Storni pier houses on cylindrical docks have very diverse communities of bivalve filter feeders, dominated by *Aulacomya atra atra* (Bala & Pastor de Ward, 2000). According to Mirto *et al.* (2000), the mussels, in general, induce changes in the sediment characteristics, leading to the oxygen penetrability inside the sediment being reduced in sites related to very dense mussels communities. Also, the biodeposition by bivalves generally provides a strong input of organic matter of high quality and availability to benthic assemblages.

Bala & Pastor de Ward (2000) indicated that the biodeposition due to the presence of *A. atra atra*

might have a local effect in the increased sedimentation next to the pier dock. They suggested that, in general, the availability of oxygen in the sediment surface layer might be due to the movement of the water column (Mirto *et al.*, 2000), with the intertidal zone next to the Admiral Storni pier being probably sheltered from the movement from the water column due to the presence of pier docks causing a higher sedimentation in this area. In addition, they argued that the biodeposition due to mussel activities could have led to a significant increase in phytopigment concentrations, and that this accumulation was apparently related to the phytoplankton blooms that were filtered and compacted by mussels and discharged to the sediments, suggesting that total phytopigments (such as chlorophyll-*a*) might be used as tracers of mussel biodeposition. In a study in Nuevo Gulf by Pastor & Bala (1996), it was shown that significant concentrations of chlorophyll-*a* near to the Admiral Storni pier were due to a biodeposition process by the presence of bivalve dominant populations of *A. atra atra*.

In contrast, the area next to Nueva beach “restinga” (sites 7-11) showed a decrease in abundance and an increase in diversity. This trend continued to Cerro Avanzado beach, where there was a decrease in the fine fraction percentage and an increase the depth of the anoxic layer. Despite the contribution of fine material together with the decomposition of this material at the disposal sites, anoxic conditions in sediment zones and changes could have been produced by the resulting strong reducing conditions that prevailed just below the sediment surface, with may inhibit settlement or survival of the most sensitive species (Boyd *et al.*, 2000).

The results of the MDS ordination and the ANOSIM tests for meiofauna data of Nuevo Gulf showed similarities, in which the structure of the benthic communities clearly differed between beaches near to and far away from the city, probably related to changes in the abiotic variables. At the sites of beaches near to the city, a higher meiofauna abundance was found next to Admiral Storni pier and the central area of Nueva beach. Moreover, a higher fine fraction percentage, a lower permeability of sediment and a thin oxygenated layer were found next to the pier. Also, at beach sites far away from the city, Kaiser, Paraná and Cerro Avanzado, a lower meiofauna abundance was found together with a lower fine fraction percentage and a thick oxygenated layer.

In general, the main factor influencing the sediment characteristics of this bay may not be the movement of the water body, which depends mainly on the instantaneous direction and intensity of the winds

and tides (Krepper & Rivas, 1979). Instead, particle size appears to be the main factor affecting the abundance and species composition of meiofauna organisms in the benthic environment (Coull, 1988). Netto *et al.* (1999a, b), at Rocas Atoll in the northeast of Brazil, showed that differences in community structure are related to meiofauna local transport processes and sedimentation caused by the intense movement of water through the atoll. According to Danovaro *et al.* (2002), in a study on the Italian coast, changes in grain size are presumably caused by alterations to the hydrodynamic conditions and the topography on a micro-scale. In this study, the sampling sites were taken at the lower intertidal level of the beach and on polychaetes communities (Maldanidae family). These community organisms construct tubes that protrude out of the sediment into the water column, and Murray *et al.* (2002) established that when these tubes were rare, local erosion is likely to occur. However, where there are abundant tubes, sediment accumulation is more probable.

The results of the data analysis showed that the average similarity was high at all sites for both groups – beaches near to or far away from the city of the Nuevo Gulf. The meiofauna data showed that turbellarians, ciliates, nematodes and polychaetes contributed most to the break-down of similarity between groups, with these taxa being responsible for the changes in community composition. Mean higher abundance of turbellarian was found at beaches near than far away from the city, which having a lower oxygenated layer and grain size. However, the benthic turbellarian systems are generally represented in oxygenated substrates and in wave-protected beaches, with little mixing of the sediments (Cannon & Faubel, 1988). In this study, mean higher abundance of ciliate was found in beaches near from the city, where have predominated the fine sediments. These results are in agreement with the reports of Corliss *et al.* (1988), where the ciliate abundances are commonly found in fine sediment (between 100 and 300 μm). Boyd *et al.* (2000) have established that nematode communities can provide a sensitive indicator of change in response to dredged material disposal. Moreover, Moreno *et al.* (2008) have shown that the nematode assemblages can increase significantly at both sites, with fine sediment and high pollution effects. In addition, the structure of the nematode assemblages in the sediment layers is also affected, probably by changes in the redox conditions caused by the bioirrigating effects (Tita *et al.*, 2000).

Nematodes are assumed to be quite resistant to sediment organic enrichment and the resulting reducing conditions (Mirto *et al.*, 2000). In this study,

although nematode was the most abundant taxon (90%) at all the sites, the highest abundance was found at sites between the Admiral Storni pier and in the central area of Nueva beach (sites 1-6), with high levels of fine fraction and low levels of both permeability and oxygen layer.

Changes in the meiofaunal structure may be induced by biodeposition effects. Nematodes together with turbellarians, among others, might represent useful indicators of biodeposition disturbance (Mirto *et al.*, 2000). Furthermore, polychaete assemblages were the most abundant in beaches far away from the city, where the sediment showed a high oxygenation. These results concur with the reports of Sutherland *et al.* (2007), where the abundance of polychaetes decreased with increasing free sulphide concentrations in the sediment. It is also important to highlight that the gastrotrich distribution showed a high abundance in two areas – the Admiral Storni pier and site 8 in Nueva beach – and was associated to the increase of sediment granulometry.

Redox potential discontinuity depth (RPD) was the abiotic factor that most influenced the meiofaunal distribution, which showed that the oxygen availability between beaches near to and far away from the city was different. This result is in agreement with the reports of Mazzola *et al.* (2000) and Mirto *et al.* (2000), who found a strong association between the meiofaunal community and the interstitial oxygen content, with the nematodes in particular revealing themselves to be more tolerant than the other organisms to sub-oxic/anoxic conditions.

CONCLUSIONS

The patterns described in the present study show that differences in the meiofauna community structure in the intertidal sandy beaches of Nuevo Gulf were significant and were related to changes in the abiotic variables, with the main factor seeming to be the oxygen availability in the sediment. Similar studies are now needed to investigate if this trend still holds over the complete spectrum of exposure along the Patagonian coast or elsewhere. In this way, it should be possible to establish general patterns of sandy beach meiofaunal variability.

RESUMEN

La composición y distribución de la meiofauna bentónica del Golfo Nuevo (Chubut, Argentina) se describen

en relación a las variables ambientales. La meiofauna y el sedimento de la zona intermareal de cuatro playas de arena con diferentes perturbaciones antrópicas fueron recolectadas en Junio de 2005. Las muestras se colectaron en 20 sitios de muestreo con tubos core de 2,5 cm de diámetro y 10 cm de profundidad. Fueron identificados 13 taxones de meiofauna, representados principalmente por nematodos, gastrotricos, ciliados y poliquetos y la abundancia de la meiofauna que fue desde $1,5 \times 10^3$ a $6,5 \times 10^3$ ind. 10 cm^{-2} . Los análisis univariado (ANOVA de una-vía) y multivariado (ANOSIM/MDS) indicaron diferencias significativas en las estructuras de la comunidad entre los sitios con efectos antrópicos y aquellos en condiciones prístinas, revelando que estas diferencias fueron entre las playas próximas y alejadas de una ciudad con la actividad portuaria. Los cambios en la estructura de la comunidad (abundancia y diversidad) pueden haber estado relacionados con gradientes ambientales próximos a la costa. El análisis BIO-ENV mostró que la profundidad de la discontinuidad del potencial redox puede ser el principal factor que influye en la distribución espacial de los organismos.

PALABRAS-CLAVES: Meiofauna; Intermareal; Sedimento arenoso; Argentina.

ACKNOWLEDGEMENTS

We very much appreciated the hospitality and excellent research facilities provided by the Centro Nacional Patagónico (CENPAT). Juan Escobar was of great help in the grain size analysis and also Gisela Pilatti by her collaboration with a stereomicroscope. We thank the Catedra of Diversidad Animal I and the Catedra de Química General of FCEfYN – UNC for providing their installations and thanks go to César, Julieta and their family for their moral support. The author C.A. Harguinteguy is a doctorate student in Biological Sciences at the National University of Córdoba and fellow at CONICET. The author thanks Dr. Paul Hobson, native speaker, for revision of the manuscript.

REFERENCES

- AUSTEN, M.C.; McEVoy, A.J. & WARWICK, R.M. 1994. The specificity of meiobenthic community responses to different pollutants: results from microcosm experiments. *Marine Pollution Bulletin*, 28:557-563.
- BALA, L.O. & PASTOR DE WARD, C.T. 2000. Biodeposition by *Aulacomya atra* Molina in patagonian coast. *Medio Ambiente*, 13:68-74.

- BARNES, N.; BAMBER, R.; MONCRIEFF, C.; SHEADER, M. & FERRERO, T. 2008. Meiofauna in closed coastal saline lagoons in the United Kingdom: Structure and biodiversity of the nematode assemblage. *Estuarine, Coastal and Shelf Science*, 79:328-340.
- BOYD, S.; REES, H. & RICHARDSON, C. 2000. Nematodes as sensitive indicators of change at dredged material disposal sites. *Estuarine, Coastal and Shelf Science*, 51:805-819.
- CANNON, L.R.G. & FAUBEL, A. 1988. Turbellaria. In: Higgins, R.P. & Thiel, H. (Eds.), *Introduction to the study of meiofauna*. Smithsonian Institution Press, Washington, D.C., London. p. 273-282.
- CHARPY, C. & CHARPY, L. 1977. *Biomasse phytoplantonique, production primaire et facteurs limitant la fertilité des eaux du Golfe "San José" (Península Valdés, Argentine)*. These de 3er Cycle, Oceanologie, Université d'Aix-Marseille II, France. 185 p.
- CHARPY-ROUBAUD, C.J.; CHARPY, L.J. & MAESTRIN, S.Y. 1982. Fertilité des eaux côtières nord-patagoniques: facteurs limitant la production du phytoplancton et potentialités d'exploitation mytilicole. *Oceanologica Acta*, 215:179-188.
- CHEN, G.; HERMAN, R. & VINCX, M. 1999. Meiofauna communities from the Straits of Magellan and the Beagle Channel. *Scientia Marina*, 63:123-132.
- CLARKE, K.; WARWICK, R.; SOMERFIELD, P. & GORLEY, R. 2005. *Methods manual: Changes in marine communities. PRIMER v6.1.10*.
- CORLISS, J.O.; HARTWIG, E. & LENK, S.E. 1988. Ciliophora. In: Higgins, R.P. & Thiel, H. (Eds.), *Introduction to the study of meiofauna*. Smithsonian Institution Press, Washington, D.C. p. 258-265.
- COULL, B.C. 1988. Ecology of the Marine Meiofauna. In: Higgins, R.P. & Thiel, H. (Eds.), *Introduction to the study of meiofauna*. Smithsonian Institution Press, Washington, D.C. p. 18-38.
- COULL, B.C. & CHANDLER, G.T. 2001. Benthos (Meiobenthos). In: Steele, J.H.; Turekian, K. & Thorpe, S.A. (Eds.), *Encyclopedia of Ocean Sciences*. Academic Press, London. p. 1705-1711.
- CROSS, R. & CURRAN, M.C. 2000. Effects of Feeding Pit Formation by Rays on an Intertidal Meiobenthic Community. *Estuarine, Coastal and Shelf Science*, 51:293-298.
- DANOVARO, R.; GAMBÌ, C.; MAZZOLA, A. & MIRTO, S. 2002. Influence of artificial reefs on the surrounding infauna: analysis of meiofauna. *ICES Journal of Marine Science: Journal du Conseil*, 59:S356-S362.
- DE TROCH, M.; VAN GANSBEKE, D. & VINCX, M. 2006. Resource availability and meiofauna in sediment of tropical seagrass beds: Local versus global trends. *Marine environmental research*, 61:59-73.
- DITLEVSEN, J. 1911. Danish free-living nematode. *Videnskabelige Meddelelser fra Dansk naturhistorisk Forening*, 63:213-256.
- DITTMANN, S. 2000. Zonation of benthic communities in a tropical tidal flat of north-east Australia. *Journal of Sea Research*, 43:33-51.
- ESTEVEZ, J.L. & DE VIDO DE MATTIO, N. 1980. *Influencia de Puerto Madryn en Bahía Nueva mediante salinidad y temperatura. Evidencia de fenómenos de Surgencia*. CONICET, Contribución. 26. 40 p.
- FLACH, E.; MUTHUMBI, A. & HEIP, C. 2002. Meiofauna and macrofauna community structure in relation to sediment composition at the Iberian margin compared to the Goban Spur (NE Atlantic). *Progress in oceanography*, 52:433-457.
- FRASCHETTI, S.; GAMBÌ, C.; GIANGRANDE, A.; MUSCO, L.; TERLIZZI, A. & DANOVARO, R. 2006. Structural and functional response of meiofauna rocky assemblages to sewage pollution. *Marine Pollution Bulletin*, 52:540-548.
- GIERE, O. 1993. *Meiobenthology*. Springer, Berlin. p. 328.
- GIERE, O.; ELEFThERIOU, A. & MURISON, D.J. 1988. Abiotic Factors. In: Higgins, R.P. & Thiel, H. (Eds.), *Introduction to the study of meiofauna*. Smithsonian Institution Press, Washington, D.C. p. 61-78.
- GOBIN, J.F. & WARWICK, R.M. 2006. Geographical variation in species diversity: A comparison of marine polychaetes and nematodes. *Journal of Experimental Marine Biology and Ecology*, 330:234-244.
- GÓMEZ NOGUERA, S. & HENDRICKX, M. 1997. Distribution and abundance of meiofauna in a subtropical coastal lagoon in the South-eastern Gulf of California, Mexico. *Marine Pollution Bulletin*, 34:582-587.
- GYEDU-ABABIO, T. & BAIRD, D. 2006. Response of meiofauna and nematode communities to increased levels of contaminants in a laboratory microcosm experiment. *Ecotoxicology and Environmental Safety*, 63:443-450.
- HALLER, M.J. 1981. *Descripción geológica de la hoja 43h, Puerto Madryn: provincia del Chubut*. Buenos Aires, Susecretaria de Esta de Minería. 45p.
- HIGGINS, R.P. & THIEL, H. 1988. *Introduction to the study of meiofauna*. Smithsonian Institution Press, Washington, D.C. 488 p.
- INGELS, J.; KIRIAKOULAKIS, K.; WOLFF, G.A. & VANREUSEL, A. 2009. Nematode diversity and its relation to the quantity and quality of sedimentary organic matter in the deep Nazaré Canyon, Western Iberian Margin. *Deep Sea Research Part I: Oceanographic Research Papers*, 56:1521-1539.
- JARAMILLO, E.; MCLACHLAN, A. & COETZEE, P. 1993. Intertidal zonation patterns of macroinfauna over a range of exposed sandy beaches in south-central Chile. *Marine Ecology Progress Series*, 101:105-116.
- KREPPER, C.M. & RIVAS, A.L. 1979. Dinámica de las aguas costeras del Golfo Nuevo. Parte 1: Medición de corrientes en superficies derivantes. *Acta Oceanográfica Argentina*, 2:83-106.
- LEDUC, D. & PROBERT, P.K. 2009. The effect of bacterivorous nematodes on detritus incorporation by macrofaunal detritivores: A study using stable isotope and fatty acid analyses. *Journal of Experimental Marine Biology and Ecology*, 371:130-139.
- LEE, M.R.; CORREA, J.A. & SEED, R. 2006. A sediment quality triad assessment of the impact of copper mine tailings disposal on the littoral sedimentary environment in the Atacama region of northern Chile. *Marine Pollution Bulletin*, 52:1389-1395.
- LEGUERRIER, D.; NIQUIL, N.; BOILEAU, N.; RZEZNIK, J.; PIERRE-GUY SAURIAU, P.-G.; LE MOINE, O. & BACHER, C. 2003. Numerical analysis of the food web of an intertidal mudflat ecosystem on the Atlantic coast of France. *Marine Ecology Progress Series*, 246:17-37.
- MAZZOLA, A.; MIRTO, S.; LA ROSA, T.; FABIANO, M. & DANOVARO, R. 2000. Fish-farming effects on benthic community structure in coastal sediments: analysis of meiofaunal recovery. *ICES Journal of Marine Science: Journal du Conseil*, 57:1454-1461.
- MIRTO, S. & DANOVARO, R. 2004. Meiofaunal colonisation on artificial substrates: a tool for biomonitoring the environmental quality on coastal marine systems. *Marine Pollution Bulletin*, 48:919-926.
- MIRTO, S.; LA ROSA, T.; DANOVARO, R. & MAZZOLA, A. 2000. Microbial and Meiofaunal Response to Intensive Mussel-Farm Biodeposition in Coastal Sediments of the Western Mediterranean. *Marine Pollution Bulletin*, 40:244-252.
- MORENO, M.; FERRERO, T.J.; GALLIZIA, I.; VEZZULLI, L.; ALBERTELLI, G. & FABIANO, M. 2008. An assessment of the spatial heterogeneity of environmental disturbance within an enclosed harbour through the analysis of meiofauna and nematode assemblages. *Estuarine, Coastal and Shelf Science*, 77:565-576.
- MURRAY, J.M.H.; MEADOWS, A. & MEADOWS, P.S. 2002. Biogeomorphological implications of microscale interactions

- between sediment geotechnics and marine benthos: a review. *Geomorphology*, 47:15-30.
- NEIRA, C.; SELLANES, J.; SOTO, A.; GUTIERREZ, D. & GALLARDO, V.A. 2001. Meiofauna and sedimentary organic matter off central Chile: response to changes caused by the 1997-1998 El Niño. *Oceanologica Acta*, 24:313-328.
- NETTO, S.; ATTRILL, M. & WARWICK, R. 1999a. The effect of a natural water-movement related disturbance on the structure of meiofauna and macrofauna communities in the intertidal sand flat of Rocas Atoll (NE, Brazil). *Journal of Sea Research*, 42:291-302.
- NETTO, S.; WARWICK, R. & ATTRILL, M. 1999b. Meiobenthic and macrobenthic community structure in carbonate sediments of Rocas Atoll (North-east, Brazil). *Estuarine, Coastal and Shelf Science*, 48:39-50.
- NETTO, S.A.; GALLUCCI, F. & FONSECA, G.F.C. 2005. Meiofauna communities of continental slope and deep-sea sites off SE Brazil. *Deep Sea Research Part I: Oceanographic Research Papers*, 52:845-859.
- NETTO, S.A.; GALLUCCI, F. & FONSECA, G.F.C. 2009. Deep-sea meiofauna response to synthetic-based drilling mud discharge off SE Brazil. *Deep Sea Research Part II: Topical Studies in Oceanography*, 56:41-49.
- PASTOR, C. & BALA, L. 1996. Estudios de base en la Bahía de Puerto Madryn (Golfo Nuevo, Chubut): Pigmentos fotosintéticos. [Baseline study in Puerto Madryn Bay (Golfo Nuevo, Chubut): Photosynthetic pigments]. *Naturalia Patagónica, Ciencias Biológicas*, 4:121-137.
- PUSCEDDU, A.; GAMBI, C.; ZEPILLI, D.; BIANCHELLI, S. & DANOVARO, R. 2009. Organic matter composition, metazoan meiofauna and nematode biodiversity in Mediterranean deep-sea sediments. *Deep Sea Research Part II: Topical Studies in Oceanography*, 56:755-762.
- RAFFAELLI, D.G. & HAWKINS, S.J. 1996. *Intertidal ecology*. Chapman & Hall, London. 356 p.
- RODRÍGUEZ, J.; LASTRA, M. & LÓPEZ, J. 2003. Meiofauna distribution along a gradient of sandy beaches in northern Spain. *Estuarine, Coastal and Shelf Science*, 58:63-69.
- SAJAN, S.; JOYDAS, T.V. & DAMODARAN, R. 2010. Meiofauna of the western continental shelf of India, Arabian Sea. *Estuarine, Coastal and Shelf Science*, 86:665-674.
- SUTHERLAND, T.F.; LEVINGS, C.D.; PETERSEN, S.A.; POON, P. & PIERCEY, B. 2007. The use of meiofauna as an indicator of benthic organic enrichment associated with salmonid aquaculture. *Marine Pollution Bulletin*, 54:1249-1261.
- TITA, G.; DESROSIERS, G.; VINCX, M. & NOZAIS, C. 2000. Predation and sediment disturbance effects of the intertidal polychaete *Nereis virens* (Sars) on associated meiofaunal assemblages. *Journal of Experimental Marine Biology and Ecology*, 243:261-282.
- VEIT-KÖHLER, G.; GERDES, D.; QUIROGA, E.; HEBBELN, D. & SELLANES, J. 2009. Metazoan meiofauna within the oxygen-minimum zone off Chile: Results of the 2001-PUCK expedition. *Deep Sea Research Part II: Topical Studies in Oceanography*, 56:1105-1111.
- WARWICK, R. & CLARKE, K. 1991. A comparison of some methods for analysing changes in benthic community structure. *Journal of the Marine Biological Association of the United Kingdom*, 71:225-244.
- WARWICK, R.M. & CLARKE, K.R. 1993. Comparing the severity of disturbance: a meta-analysis of marine macrobenthic community data. *Marine ecology progress series. Oldendorf*, 92:221-231.
- DE WIT, R.; STAL, L.J.; LOMSTEIN, B.A.; HERBERT, R.A.; VAN GEMERDEN, H.; VIAROLI, P.; CECHESELLI, V.U.; RODRÍGUEZ-VALERA, F.; BARTOLI, M. & GIORDANI, G. 2001. ROBUST: the role of buffering capacities in stabilising coastal lagoon ecosystems. *Continental Shelf Research*, 21:2021-2041.
- YAMANAKA, T.; RAFFAELLI, D. & WHITE, P.C.L. 2010. Physical determinants of intertidal communities on dissipative beaches: Implications of sea-level rise. *Estuarine, Coastal and Shelf Science*, 88:267-278.

Aceito em: 01.11.2012

Impresso em: 20.12.2012

Apoio:



Ministério
da Educação

Ministério da
Ciência e Tecnologia



EDITORIAL COMMITTEE

Publisher: Museu de Zoologia da Universidade de São Paulo. Avenida Nazaré, 481, Ipiranga, CEP 04263-000, São Paulo, SP, Brasil.

Editor-in-Chief: Carlos José Einicker Lamas, Serviço de Invertebrados, Museu de Zoologia, Universidade de São Paulo, Caixa Postal 42.494, CEP 04218-970, São Paulo, SP, Brasil. E-mail: editormz@usp.br.

Associate Editors: Mário César Cardoso de Pinna (*Museu de Zoologia, Universidade de São Paulo, Brasil*); Luís Fábio Silveira (*Museu de Zoologia, Universidade de São Paulo, Brasil*); Marcos Domingos Siqueira Tavares (*Museu de Zoologia, Universidade de São Paulo, Brasil*); Sérgio Antonio Vanin (*Instituto de Biociências, Universidade de São Paulo, Brasil*); Hussam El Dine Zaher (*Museu de Zoologia, Universidade de São Paulo, Brasil*).

Editorial Board: Rüdiger Bieler (*Field Museum of Natural History, U.S.A.*); Walter Antonio Pereira Boeger (*Universidade Federal do Paraná, Brasil*); Carlos Roberto Ferreira Brandão

(*Universidade de São Paulo, Brasil*); James M. Carpenter (*American Museum of Natural History, U.S.A.*); Ricardo Macedo Corrêa e Castro (*Universidade de São Paulo, Brasil*); Mario de Vivo (*Universidade de São Paulo, Brasil*); Marcos André Raposo Ferreira (*Museu Nacional, Rio de Janeiro, Brasil*); Darrel R. Frost (*American Museum of Natural History, U.S.A.*); William R. Heyer (*National Museum of Natural History, U.S.A.*); Ralph W. Holzenthal (*University of Minnesota, U.S.A.*); Adriano Brilhante Kury (*Museu Nacional, Rio de Janeiro, Brasil*); Gerardo Lamas (*Museo de Historia Natural "Javier Prado", Lima, Peru*); John G. Maisey (*American Museum of Natural History, U.S.A.*); Naércio Aquino Menezes (*Universidade de São Paulo, Brasil*); Christian de Muizon (*Muséum National d'Histoire Naturelle, Paris, France*); Nelson Papavero (*Universidade de São Paulo, Brasil*); James L. Patton (*University of California, Berkeley, U.S.A.*); Richard O. Prum (*University of Kansas, U.S.A.*); Olivier Rieppel (*Field Museum of Natural History, U.S.A.*); Miguel Trefaut Urbano Rodrigues (*Universidade de São Paulo, Brasil*); Randall T. Schuh (*American Museum of Natural History, U.S.A.*); Ubirajara Ribeiro Martins de Souza (*Universidade de São Paulo, Brasil*); Paulo Emilio Vanzolini (*Universidade de São Paulo, Brasil*); Richard P. Vari (*National Museum of Natural History, U.S.A.*).

INSTRUCTIONS TO AUTHORS - (April 2007)

General Information: *Papéis Avulsos de Zoologia* (PAZ) and *Arquivos de Zoologia* (AZ) cover primarily the fields of Zoology, publishing original contributions in systematics, paleontology, evolutionary biology, ontogeny, faunistic studies, and biogeography. *Papéis Avulsos de Zoologia* and *Arquivos de Zoologia* also encourage submission of theoretical and empirical studies that explore principles and methods of systematics.

All contributions must follow the International Code of Zoological Nomenclature. Relevant specimens should be properly curated and deposited in a recognized public or private, non-profit institution. Tissue samples should be referred to their voucher specimens and all nucleotide sequence data (aligned as well as unaligned) should be submitted to GenBank (www.ncbi.nih.gov/genbank) or EMBL (www.ebi.ac.uk).

Peer Review: All submissions to *Papéis Avulsos de Zoologia* and *Arquivos de Zoologia* are subject to review by at least two referees and the Editor-in-Chief. All authors will be notified of submission date. Authors may suggest potential reviewers. Communications regarding acceptance or rejection of manuscripts are made through electronic correspondence with the first or corresponding author only. Once a manuscript is accepted providing changes suggested by the referees, the author is requested to return a revised version incorporating those changes (or a detailed explanation of why reviewer's suggestions were not followed) within fifteen days upon receiving the communication by the editor.

Proofs: Page-proofs with the revised version will be sent to e-mail the first or corresponding author. Page-proofs *must be returned to the editor, preferentially within 48 hours*. Failure to return the proof promptly may be interpreted as approval with no changes and/or may delay publication. Only necessary corrections in proof will be permitted. Once page proof is sent to the author, further alterations and/or significant additions of text are permitted only at the author's expense or in the form of a brief appendix (note added in proof).

Submission of Manuscripts: Manuscripts should be sent to the **SciELO Submission** (<http://submission.scielo.br/index.php/paz/login>), along with a submission letter explaining the importance and originality of the study. Address and e-mail of the corresponding author must be always updated since it will be used to send the 50 reprints in titled by the authors. Figures, tables and graphics **should not** be inserted in the text. Figures and graphics should be sent in separate files with the following formats: ".JPG" and ".TIF" for figures, and ".XLS" and ".CDR" for graphics, with 300 DPI of minimum resolution. Tables should be placed at the end of the manuscript.

Manuscripts are considered on the understanding that they have not been published or will not appear elsewhere in substantially the same or abbreviated form. The criteria for acceptance of articles are: quality and relevance of research, clarity of text, and compliance with the guidelines for manuscript preparation.

Manuscripts should be written preferentially in English, but texts in Portuguese or Spanish will also be considered. Studies with a broad coverage are encouraged to be submitted in English. All manuscripts should include an abstract and key-words in English and a second abstract and key-words in Portuguese or Spanish.

Authors are requested to pay attention to the instructions concerning the preparation of the manuscripts. Close adherence to the guidelines will expedite processing of the manuscript.

Manuscript Form: Manuscripts should not exceed 150 pages of double-spaced, justified text, with size 12 and source Times New Roman (except for symbols). Page format should be A4 (21 by 29.7 cm), with 3 cm of margins. The pages of the manuscript should be numbered consecutively.

The text should be arranged in the following order: **Title Page, Abstracts with Key-Words, Body of Text, Literature Cited, Tables, Appendices, and Figure Captions**. Each of these sections should begin on a new page.

(1) **Title Page:** This should include the **Title, Short Title, Author(s) Name(s)** and **Institutions**. The title should be concise and, where appropriate, should include mention of families and/or higher taxa. Names of new taxa should not be included in titles.

(2) **Abstract:** All papers should have an abstract in **English** and another in **Portuguese or Spanish**. The abstract is of great importance as it may be reproduced elsewhere. It should be in a form intelligible if published alone and should summarize the main facts, ideas, and conclusions of the article. Telegraphic abstracts are strongly discouraged. Include all new taxonomic names for referencing purposes. Abbreviations should be avoided. It should not include references. Abstracts and key-words should not exceed 350 and 5 words, respectively.

(3) **Body of Text:** The main body of the text should include the following sections: **Introduction, Material and Methods, Results, Discussion, Conclusion, Acknowledgments, and References at end**. Primary headings in the text should be in capital letters, in bold and centered. Secondary headings should be in capital and lower case letters, in bold and centered. Tertiary headings should be in capital and lower case letters, in bold and indented at left. In all the cases the text should begin in the following line.

(4) **Literature Cited:** Citations in the text should be given as: Silva (1998) *or* Silva (1998:14-20) *or* Silva (1998: figs. 1, 2) *or* Silva (1998a, b) *or* Silva & Oliveira (1998) *or* (Silva, 1998) *or* (Rangel, 1890; Silva & Oliveira, 1998a, b; Adams, 2000) *or* (Silva, *pers. com.*) *or* (Silva *et al.*, 1998), the latter when the paper has three or more authors. The reference need not be cited when authors and date are given only as authority for a taxonomic name.

(5) **References:** The literature cited should be arranged strictly alphabetically and given in the following format:

- **Journal Article** - Author(s). Year. Article title. *Journal name*, volume: initial page-final page. Names of journals must be spelled out in full.
- **Books** - Author(s). Year. *Book title*. Publisher, Place.
- **Chapters of Books** - Author(s). Year. Chapter title. *In*: Author(s) ou Editor(s), *Book title*. Publisher, Place, volume, initial page-final page.
- **Dissertations and Theses** - Author(s). Year. *Dissertation title*. (Ph.D. Dissertation). University, Place.
- **Electronic Publications** - Author(s). Year. *Title*. Available at: <electronic address>. Access in: date.

Tables: All tables must be numbered in the same sequence in which they appear in text. Authors are encouraged to indicate where the tables should be placed in the text. They should be comprehensible without reference to the text. Tables should be formatted with vertical (portrait), not horizontal (landscape), rules. In the text, tables should be referred as Table 1, Tables 2 and 4, Tables 2-6. Use "TABLE" in the table heading.

Illustrations: Figures should be numbered consecutively, in the same sequence that they appear in the text. Each illustration of a composite figure should be identified by capital letters and referred in the text as: Fig. 1A, Fig. 1B, for example. When possible, letters should be placed in the left lower corner of each illustration of a composite figure. Hand-written lettering on illustrations is unacceptable. Figures should be mounted in order to minimize blank areas between each illustration. Black and white or color photographs should be digitized in high resolution (300 DPI at least). Use "Fig(s)," for referring to figures in the text, but "FIGURE(S)" in the figure captions and "fig(s)," when referring to figures in another paper.

Responsability: Scientific content and opinions expressed in this publication are sole responsibility of the respective authors.
Copyrights: The journals *Papéis Avulsos de Zoologia* and *Arquivos de Zoologia* are licensed under a Creative Commons Licence (<http://creativecommons.org>).

For other details of manuscript preparation of format, consult the CBE Style Manual, available from the Council of Science Editors (www.councilscienceeditors.org/publications/style).
Papéis Avulsos de Zoologia and *Arquivos de Zoologia* are publications of the Museu de Zoologia da Universidade de São Paulo (www.mz.usp.br).
Always consult the Instructions to Authors printed in the last issue or in the electronic home pages: www.scielo.br/paz or www.mz.usp.br/publicacoes.